## Stock Assessment of hoki (Macruronus magellanicus) in the Falkland Islands



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## Summary

Commercial catches of hoki Macruronus magellanicus in Falkland Islands licenced fisheries were 7,629 tonnes ( $t$ ) in 2020, the eighth lowest catch since 1990. Nearly $97 \%$ of the commercial catches were by finfish licences (G-, and W-). Fishing effort decreased steeply since the early 2000's; as a result, commercial CPUE in 2020 was the second highest level ( $1,018 \mathrm{~kg} / \mathrm{h}$ ) since 1990. In contrast, CPUE from summer (February) surveys had its third lowest level ( $75 \mathrm{~kg} / \mathrm{h}$ ) in 2020 since 2010.

Following recommendations of the MacAlister Elliott \& Partners external review, Total Allowable Catch (TAC) was calculated according to the ICES category 5 assessment framework: three-year catch average. The hoki TAC for 2022 is recommended at $6,478 \mathrm{t}$, which represents a decrease of $15 \%$ from the total commercial catch in $2020(7,629 \mathrm{t})$.

Length-based indicators were scored on a traffic-light scale of green, yellow, red. Conservation of immature individuals and large individuals was of concern (yellow) or negative (red), and conservation of mega-spawners was negative since 2002. Optimal yield was of concern or negative, mainly for females. Maximum sustainable yield was negative for females, except from 2006 to 2015. Maximum sustainable yield fluctuated for males but was negative since 2017. Asymptotic lengths ( $L_{\infty}$ ) increased significantly over the past decade for males. However, lengths at 50\% maturity decreased since 2011 for females and since 2002 for males. Individuals $>25 \mathrm{~cm}$ pre-anal length size classes and $\geq 4$ years old were dominant early in the time series (i.e., 2002-2003, and 2006-2007). Individuals < 25 cm pre-anal length and < 4 years old were more common in the following years, with a peak of nearly 15 cm preanal length and 1-year old individuals observed in 2017. Recruitment to the fishery of 1-and 2-years old individuals was detected every 2 to 4 years.

## Introduction

Hoki Macruronus magellanicus Lönnberg, 1907 (Merlucciidae) is a highly migratory pelagic-demersal fish that inhabits 30-500 m depth (Froese \& Pauly 2021). This species occurs in temperate shelf and slope waters of the Southeast Pacific from $29^{\circ} \mathrm{S}$ (southern Chile) and of the Southwest Atlantic from $33^{\circ} \mathrm{S}$ to $57^{\circ} \mathrm{S}$ around Cape Horn, including Argentina and Falkland Islands (Wöhler \& Giussi 2001; Schuchert et al. 2010; Froese \& Pauly 2021). Hoki is one of the most abundant species on the Patagonian shelf; however, it is not highly abundant in Falkland Islands waters as the Falkland Island Conservation Zone is located at the edge of its distribution (Falkland Islands Government 2021). Hoki in the Southwest Atlantic and in the Southeast Pacific belong to the same population (McKeown et al. 2015), via migrations around Cape Horn and throughout the channels of Tierra del Fuego (Wöhler \& Giussi 2001). In agreement with these findings, genetic studies found that individuals from the Argentine coast, and from near the west $\left(52^{\circ} \mathrm{S}, 64^{\circ} \mathrm{W}\right)$ and southwest $\left(54^{\circ} 18^{\prime} \mathrm{S}, 64^{\circ} 43^{\prime} \mathrm{W}\right)$ edge of the Falkland Island Conservation Zone belong to the same population (D'Amato \& Carvalho 2005; D'Amato 2006). Therefore, hoki from the Falkland Islands, Argentina and Chile will be considered as a single stock for the purpose of this report.

The main spawning aggregations have been encountered in the vicinity of Guamin Island, Chile, between $43^{\circ} \mathrm{S}$ and $48^{\circ} \mathrm{S}$ (Payá et al. 2002). Smaller aggregations of spawning fish and juveniles have also been found in the Southwest Atlantic in the Gulf of San Matias and in the Gulf of San Jorge in Argentina (Wöhler \& Giussi 2001), and on the shelf edge east of the Falkland Islands (Giussi 1996). Larvae are present on either side of the Magellanic Strait ( $53^{\circ} \mathrm{S}$ ), near Cape Horn ( $55^{\circ} \mathrm{S}$ ), and farther north in coastal areas of the Atlantic Ocean (Niklitschek et al. 2014). After winter spawning, part of the hoki population migrates in spring to feeding grounds in the slope areas of the Falkland Current Front (west of the Falkland Islands) (Brickle et al. 2009; Arkhipkin et al. 2012), and in summer it mainly occupies the warmer northern Falkland Islands' shelf (Brickle et al. 2009).

## Methods

## ICES advice rules

In 2020, hoki was included in a Falkland Islands Government finfish stock assessment and management review (MEP 2020). The MEP report recommended stock assessments for most commercial finfish species to be based on the ICES advice rules (ICES 2012, 2018a),
referencing applicable categories of data availability and quality. MEP (2020) recommended using a category 5 assessment framework for hoki. MEP (2020) also recommended exploring ancillary stock status information from ICES data limited methods such as length-based indicators. Therefore, a Length-Based Indicator method (LBI) (ICES 2015) was used to provide a suite of indicators based on combinations of catch-at-size distributions and life-history parameters.

## Commercial fishery data

Commercial fishing around the Falkland Islands was not distinguished from other parts of the Southwest Atlantic prior to 1982 and catch data by species were recorded systematically from 1987 only (Falkland Islands Government 1989). Therefore, total hoki catch data were examined from 1987 to 2020 (Falkland Islands [http://www.fig.gov.fk/fisheries/publications/fishery-statistics; Falkland Islands Government 2021], Argentina [https://www.agroindustria.gob.ar/sitio/areas/pesca maritima/desembarques/; Sánchez et al. 2012; Navarro et al. 2014, 2019], and Chile [http://www.sernapesca.cl/informes/estadisticas; SERNAPESCA 1990, 2000, 2011, 2021]). LOESS (span $=0.75$, degree $=2$ ) was implemented to examine the pattern of the association between Falkland Islands and Argentine, and between Falkland Islands and Chilean commercial annual catches of hoki from 1987 to 2020.

Commercial catches of hoki in Falkland Islands waters were examined by licence type for 2020. Exploratory analysis showed that finfish G- and W-licences contributed most hoki catch and catch-per-unit-effort (CPUE) across years. Therefore, spatial distribution of the 2020 monthly CPUE average was estimated from G- and W-licences, excluding the finfish Alicence. CPUE was also estimated per year and per month from G - and W -licences only.

## Scientific surveys data

Biomass estimates and the spatial distribution of hoki were examined from austral summer scientific surveys (groundfish and D. gahi pre-season surveys) carried out in February 2010, 2011, and 2015 - 2021 in Falkland Islands waters (Ramos \& Winter 2021). A trend of the biomass time series from 2010 to 2021 was calculated using LOESS (span $=0.75$, degree
$=2$ ). Biomass ratios between the most recent February surveys (2021) and the first February surveys (2010) were estimated as a proxy of the change in biomass over time. Significance of difference and $95 \%$ confidence intervals of the change in biomass were computed from the randomized re-samples of the survey biomass estimates (Ramos \& Winter 2021). Biomass estimates, the spatial distribution of hoki, and biomass ratios were also examined (following Ramos \& Winter 2021) from scientific surveys carried out in austral winter, during July 2017 (Gras et al. 2017; Winter et al. 2017) and July 2020 (Randhawa et al. 2020; Winter et al. 2020).

## ICES Category 5 Total Allowable Catch

The category 5 assessment framework is based on the average catches from the 3 previous years (MEP 2020). Therefore, Total Allowable Catch (TAC) for the year 2022 was estimated based on the in-zone average catch from 2018 to 2020, excluding experimental (Elicence) and out-of-zone catches (O-licence), whereby no hoki catches were reported out-ofzone during those years:

$$
T A C_{-} 5_{2022}=\overline{C_{2018} \text { to } 2020}
$$

Where C = Catch ( t ).

## Length analyses

## Length Based Indicators

ICES $(2015,2018 b)$ recommends the LBI method, which provides a suite of indicators based on combinations of catch-at-size distributions, life-history parameters such as $L_{\infty}$ (Haddon 2001) and L50 (length at 50\% maturity; Cope \& Punt 2009). L $\infty$ and L50 parameters were assessed for females and males separately.

LBI method was applied to all years from which observer length measurements of hoki were available and reported as random samples (FIFD database codes $R$ and S), i.e., years 2002 to 2020. Pre-anal lengths of up to one hundred individuals were measured to the lowest centimetre per trawl. Because finfish trawls are restricted to larger meshes than calamari trawls, only observer length measurements taken in finfish-licensed fisheries were used, to avoid biasing length-frequency distributions if proportionally more samples are recorded
from one fishery or another in different years. Skate and Illex trawls were also excluded; while skate and Illex currently do not have different mesh allowances from finfish, their different targets could also relate to characteristically different length-frequency distributions of hoki.

The procedure for identifying finfish-licensed observer samples is described in Appendix I. LBI method indicators were then selected and scored using Tables 2.1.1.4.1 and

### 2.1.2.2 in ICES (2015) as templates:

1) Length at half the modal catch length should be bigger than L50, for conservation of immature fish ( $L_{c} / L 50>1$ ). Note that length at half the modal catch length may be poorly defined if the catch length-frequency distribution is not smooth and unimodal.
2) Length at cumulative $25^{\text {th }}$ percentile of catch numbers should be bigger than L50, for conservation of immature fish ( $\mathrm{L}_{25 \%} / \mathrm{L} 50>1$ ).
3) Mean length of the largest $5 \%$ of individuals in the catch should be at least $80 \%$ of the asymptotic length, as a benchmark that enough large individuals are in the stock ( $L_{m a x 5 \%}$ / $\mathrm{L}_{\infty}>0.8$ ).
4) 'Mega-spawners' should comprise at least $30 \%$ of the catch (thus implicitly represent at least $30 \%$ of the stock), as large, old fish disproportionately benefit the resilience of the population (Froese 2004) ( $\mathrm{P}_{\text {mega }}>0.3$ ). Mega-spawners are defined as individuals larger than optimum length (Lopt) $+10 \%$, where Lopt is described as the length at which growth rate is maximum (ICES 2015), or the length at which total biomass of a year-class reaches its maximum value (Froese \& Binohlan 2000). Lopt $=3 \cdot \mathrm{~L}_{\infty} \cdot\left(3+\mathrm{Mk}^{-1}\right)^{-1}$ (Beverton 1992), where M is instantaneous natural mortality, k is the rate of curvature of the von Bertalanffy growth function, and the ratio $\mathrm{Mk}^{-1}$ is set in WKLIFE V software (ICES 2015) at the standard constant of 1.5 (Jensen 1996).
5) Mean length of individuals larger than $L_{c}$ ( $L_{\text {meanc }}$ ) should be approximately equal to Lopt, for optimal yield ( $L_{\text {meanc }} / L_{\text {opt }} \approx 1$ ).
6) $L_{\text {meanc }}$ should be equal or bigger to the length-based proxy for MSY ( $L_{F=M}$ ), for producing maximum sustainable yield ( $L_{\text {meanc }} / L_{F=M} \geq 1$ ). $L_{F=M}$ implements the premise that MSY is attained when fishing mortality equals natural mortality (Froese et al. 2018), and in WKLIFE V software (ICES 2015) is computed as ( $3 \cdot \mathrm{~L}_{\mathrm{c}}+\mathrm{L}_{\infty}$ )/4.

Margins of variability of the six indicators were estimated by randomly re-sampling $30,000 \times$ on the normal distribution each year's fits of $L_{\infty}$ and $L 50$ to the LOESS smooths.

Indicators were scored against the 'traffic light' scale (ICES 2015) with reference criteria >1 for conservation of immature fish, $>0.8$ for conservation of large fish, and $>0.3$ for conservation of mega-spawners. The score was green if the lower $95 \%$ quantile of the resampled iterations was $>1,>0.8$, and $>0.3$, yellow if $1,0.8$, and 0.3 were between the lower and upper $95 \%$ quantiles, and red if the upper $95 \%$ quantile of the re-sampled iterations was $<1,<0.8$, and $<0.3$. The use of the margins of variability means that same empirical values of indicators may be scored different colours in different years. Reference criterion $\approx 1$ for optimal yield was green if the lower and upper $95 \%$ quantiles spanned 1.0, yellow if the lower and upper $95 \%$ quantiles spanned 0.9 (the threshold used in ICES 2015) without spanning 1.0, and red otherwise. Reference criterion $\geq 1$ for MSY was scored the same as $>1$, except that empirical values $\geq 1$ were automatically green.

## Length-age relationship

$L_{\infty}$ was calculated from the von Bertalanffy growth function, modelled to hoki length and age data from the FIFD database with nonlinear least-squares fitting using the R package 'fishmethods' (Nelson 2019). Hoki length and age data were available for years 1988-2018, with status of age data advised 'with caution' (Lee et al. 2020) as verification of these ages is in progress. Variability of $L_{\infty}$ and the other von Bertalanffy parameters was estimated by bootstrapping. Residuals of the von Bertalanffy model fit were randomly re-sampled with replacement, added back to the expected lengths; these newly generated data were re-fit to the von Bertalanffy function, and the $95 \%$ quantiles of 30,000 iterations retained as confidence intervals. Inter-annual trends of the von Bertalanffy parameters were calculated by LOESS (span $=0.90$, degree $=2$, weighted by inverse variance), and the LOESS smooth fits applied to the LBI indicators to mitigate unevenness over the time series.

## Length at 50\% maturity (L50)

Length at $50 \%$ maturity (L50) was calculated as the mid-point of the binomial logistic regression of maturity vs. length (Heino et al. 2002). Gonadal maturity is cyclical as fish pass through pre- to post-spawning phases, and definitive maturity assignments can only be made that stages 1 are immature and stages 3 or higher are mature (H. Randhawa, FIFD, personal communication). Therefore, maturity assignment was simplified to a dichotomous classification of juvenile ( $0-1$ ) or adult ( $3+$ ), omitting stage 2 (Winter 2018). Hoki maturities
were available from all years 2002 to 2020. The aggregates of L50 were plotted against years and trends calculated with LOESS smooths (span $=0.90$, degree $=2$ ), also weighted by inverse variance of each year's binomial logistic regression. These LOESS smooth fits were also used for LBI parameterization per year.

## Length frequencies

Length frequencies were examined yearly for females and males separate to describe patterns in length from 2002 to 2020. Lengths of individuals sampled randomly (FIFD database codes R and S ) on finfish bottom trawl vessels, i.e., G -, and W - licences, were included in the analysis. Juveniles and unsexed individuals were excluded from the analysis.

## Results

## ICES advice rules

## Commercial fishery data

Hoki catches in Falkland Islands waters have averaged 4,129 t per year since 1987, representing approximately 9\% of the Falkland Islands, Argentine, and Chilean combined annual catch (Fig. 1; Appendix II).


Fig. 1. Annual commercial catch of hoki in Falkland Islands, Argentine and Chilean waters. Falkland Islands commercial catch data exclude experimental (E-licence) and out-of-zone (Olicence) licences.

Falkland Islands and Argentine annual hoki catches were significantly positively associated when Argentine catches were approximately between $30,000 \mathrm{t}$ and $95,000 \mathrm{t}$. Falkland Islands and Chilean annual hoki catches were significantly positively associated for Chilean catches up to approximately 100,000 t (Fig. 2).



Fig. 2. Annual commercial catches of hoki, Falkland Islands vs. Argentina (left) and Falkland Islands vs. Chile (right), from 1987 to 2020. Blue lines: LOESS smooths $\pm 95 \%$ confidence intervals.

During 2020 a total of 7,643 t of hoki were reported caught in Falkland Islands waters, of which 7,629 t were reported under commercial licences, i.e., excluding the experimental E-licence. Two finfish licences alone (W- and G-licences) accounted for $97 \%$ of the total hoki catch (Table I).

Table I. Catches by licence of hoki in Falkland Islands waters during 2020.

| Licence | Target species | Catch (t) | Catch (\%) |
| :--- | :--- | ---: | ---: |
| W | Restricted finfish | 5938.12 | 77.69 |
| G | Restricted finfish and IIlex | 1445.98 | 18.92 |
| A | Unrestricted finfish | 128.38 | 1.68 |
| C | Calamari $^{\text {st }}$ season | 87.63 | 1.15 |
| X | Calamari $^{\text {nd }}$ season | 29.20 | 0.38 |
| E | Experimental | 13.84 | 0.18 |
| B | IIlex squid | 0.15 | 0.00 |
| F | Skates and rays | 0.00 | 0.00 |
| L | Toothfish (longline) | 0.00 | 0.00 |
| S | Southern blue whiting and hoki | 0.00 | 0.00 |
| O | Outside Falkland Islands waters | 0.00 | 0.00 |
| Total |  | $7,643.30$ | 100.00 |

W- and G-licence CPUE had an increasing trend in the time series; the highest CPUE was recorded in $2019(1,211 \mathrm{~kg} / \mathrm{h})$ and the second highest CPUE occurred in $2020(1,018 \mathrm{~kg} / \mathrm{h})$ (Fig. 3).


Fig. 3. Yearly catch, effort, and CPUE of hoki in Falkland Islands waters, calculated from Gand W -licensed vessels.

Average monthly CPUE since 1990 had a declining trend from January through December, with the highest values recorded in February ( $729 \mathrm{~kg} / \mathrm{h}$ ), March ( $737 \mathrm{~kg} / \mathrm{h}$ ), and April ( $684 \mathrm{~kg} / \mathrm{h}$ ). Secondary peaks were observed in June and in October; the peak in June at > 200 m depth represents the subadult slope foraging immigration and at $<200 \mathrm{~m}$ depth represents the adult shelf spawning emigration (Laptikhovsky 2007). In 2020, the highest CPUEs were recorded in January and February ( $1,798 \mathrm{~kg} / \mathrm{h}$ and $1,427 \mathrm{~kg} / \mathrm{h}$, respectively). Secondary peaks were also evident in June ( $994 \mathrm{~kg} / \mathrm{h}$ ) and in October (1,008 kg/h) (Fig. 4;

Appendix III). Hoki were caught mainly to the west and southwest of West Falkland, between $51^{\circ} \mathrm{S}$ and $53^{\circ} \mathrm{S}$, and between $61^{\circ} \mathrm{W}$ and $63.5^{\circ} \mathrm{W}$ mainly during the first half of 2020; minor catches were also reported to the north in the FICZ (Appendix IV).


Fig. 4. Average monthly catch, effort, and CPUE of hoki in Falkland Islands waters for 2020 (dark blue line), and average since 1990 (light blue line), calculated from G- and W-licensed vessels.

## Scientific surveys data

## Summer surveys (February)

The biomass of hoki during the 2021 February surveys ( $312,118 \mathrm{t}$ ) was $1.1 \times$ the biomass of the 2010 February surveys ( 272,080 t; Table II; Fig. 5). However, only 14,376 out of 30,000 paired re-samples had higher biomass estimate values in February 2021 than in

February 2010 (47.9\%), therefore not significant at $p>0.05$. During the February surveys, hoki were dispersed through the FICZ in 2010, 2011, and 2015. From 2016, hoki were mainly aggregated to the southwest edge of the FICZ (Appendix V).

Table II. Summer (February) surveys catch and effort, and biomass estimates (mean $\pm 95 \%$ confidence intervals) of hoki in Falkland Islands waters.

| Year | Survey | Trawls <br> (n) | Swept area ( $\mathrm{km}^{2}$ ) | Effort <br> (h) | Catch (kg) | $\begin{aligned} & \text { CPUE } \\ & \text { (kg/h) } \end{aligned}$ | Biomass <br> ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Groundfish | 87 | 17.04 | 87.52 | 49656.01 | 567.39 | $\begin{gathered} 272080.22 \\ (197644.96-472481.97) \end{gathered}$ |
|  | D. gahi | 55 | 42.29 | 109.27 | 30124.00 | 275.69 |  |
|  | Total | 142 | 59.33 | 196.79 | 79780.01 | 405.42 |  |
| 2011 | Groundfish | 88 | 17.21 | 88.00 | 28405.39 | 322.79 | $\begin{gathered} 225981.56 \\ (173396.03-287362.59) \end{gathered}$ |
|  | D. gahi | 58 | 40.04 | 110.63 | 27594.30 | 249.42 |  |
|  | Total | 146 | 57.25 | 198.63 | 55999.69 | 281.92 |  |
| 2015 | Groundfish | 89 | 16.72 | 90.17 | 9768.23 | 108.34 | $\begin{gathered} 129562.42(40753.69- \\ 175529.10) \end{gathered}$ |
|  | D. gahi | 57 | 46.90 | 111.50 | 16596.00 | 148.84 |  |
|  | Total | 146 | 63.61 | 201.67 | 26364.23 | 130.73 |  |
| 2016 | Groundfish | 90 | 17.64 | 91.42 | 21666.57 | 237.01 | $\begin{gathered} 167312.12 \\ (83510.52-231697.65) \end{gathered}$ |
|  | D. gahi | 56 | 54.46 | 107.92 | 17248.42 | 159.83 |  |
|  | Total | 146 | 72.10 | 199.33 | 38914.99 | 195.23 |  |
| 2017 | Groundfish | 90 | 18.52 | 92.00 | 3206.21 | 34.85 | $\begin{gathered} 28863.12 \\ (16842.07-39751.29) \end{gathered}$ |
|  | D. gahi | 58 | 54.09 | 117.00 | 488.32 | 4.17 |  |
|  | Total | 148 | 72.62 | 209.00 | 3694.53 | 17.68 |  |
| 2018 | Groundfish ${ }^{\text {a }}$ | 97 | 20.47 | 96.42 | 29334.80 | 304.25 | $\begin{gathered} 139665.90 \\ (91380.06-203699.81) \end{gathered}$ |
|  | D. gahi | 59 | 36.87 | 100.83 | 682.10 | 6.76 |  |
|  | Total | 156 | 57.35 | 197.25 | 30016.90 | 152.18 |  |
| 2019 | Groundfish | 79 | 17.22 | 79.00 | 7315.40 | 92.60 | $\begin{gathered} 41346.89 \\ (6569.34-188598.04) \end{gathered}$ |
|  | D. gahi | 52 | 72.70 | 97.05 | 238.50 | 2.46 |  |
|  | Total | 131 | 89.93 | 176.05 | 7553.90 | 42.91 |  |
| 2020 | Groundfish ${ }^{\text {a }}$ | 80 | 17.04 | 79.95 | 14323.13 | 179.15 | $\begin{gathered} 77727.54 \\ (20133.68-165424.57) \end{gathered}$ |
|  | D. gahi | 59 | 86.80 | 112.52 | 59.15 | 0.53 |  |
|  | Total | 139 | 103.84 | 192.47 | 14382.28 | 74.73 |  |
| 2021 | Groundfish | 80 | 16.34 | 79.48 | 30457.98 | 383.20 | $\begin{gathered} 312118.42 \\ (93792.22-737156.05) \end{gathered}$ |
|  | D. gahi | 55 | 90.64 | 111.22 | 373.83 | 3.36 |  |
|  | Total | 135 | 106.99 | 190.70 | 30831.81 | 161.68 |  |

[^0]

Fig. 5. Hoki biomass estimates (red) and smoothed biomass trend (LOESS; span $=0.75$, degree $=2$ ) from summer (February) surveys in Falkland Islands waters. The dark blue line and the light blue area are the LOESS smooth $\pm 95 \%$ confidence intervals.

## Winter surveys (July)

The estimated biomass of hoki in the July 2020 survey ( $41,553 \mathrm{t}$ ) was $38 \%$ of the biomass estimated in the July 2017 survey (108,207; Table III). However, a total of 27,661 out of 30,000 paired re-samples had higher biomass estimate values in July 2017 than in July 2020 ( $92.2 \%$ ), thus not significant at $p>0.05$. In July 2017, aggregations of hoki were detected to the east and west in the Falkland Islands Conservation Zones, whereas in July 2020 hoki were mainly aggregated to the southwest (Appendix VI). Differences in biomass estimates between February and July surveys are likely due to the migratory pattern of hoki.

Table III. Winter (July) surveys catch and effort, and biomass estimates (mean $\pm 95 \%$ confidence intervals) of hoki in Falkland Islands waters.

| Year | Survey | Trawls <br> (n) | Swept area ( $\mathrm{km}^{2}$ ) | Effort <br> (h) | Catch (kg) | $\begin{aligned} & \text { CPUE } \\ & (\mathrm{kg} / \mathrm{h}) \end{aligned}$ | Biomass <br> ( t ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | Groundfish | 74 | 15.40 | 74 | 6450.40 | 87.17 | $\begin{gathered} 108206.57 \\ (27528.30-182684.71) \end{gathered}$ |
|  | D. gahi ${ }^{\text {a }}$ | 59 | 54.70 | 114 | 108267.50 | 949.71 |  |
|  | Total | 133 | 70.10 | 188 | 114717.90 | 610.20 |  |
| 2020 | Groundfish ${ }^{\text {b }}$ | 33 | 7.14 | 33 | 1721.86 | 52.15 | $\begin{gathered} 41552.76 \\ (8229.11-65265.20) \end{gathered}$ |
|  | D. gahi | 55 | 98.57 | 101 | 232.34 | 2.29 |  |
|  | Total | 88 | 105.71 | 134 | 1954.20 | 14.55 |  |

${ }^{\text {a }}$ An additional one-day transect of four trawls was taken in shallow inshore waters to sample for juvenile toothfish. These four trawls were not included in analyses as their locations were not relevant to the distribution of hoki.
${ }^{\text {b }}$ Twelve additional trawls were conducted in high seas during the July 2020 survey; these trawls were not included in the analyses.

## ICES Category 5 Total Allowable Catch

Total Allowable Catch (TAC) for the year 2022 under the ICES category 5 assessment framework was set at $6,478 \mathrm{t}$, as follows:

$$
T A C_{-} 5_{2022}=\frac{4408+7398+7629}{3}=\mathbf{6 4 7 8}
$$

Where the in-zone catch for 2018 (4,408 t), 2019 ( $7,398 \mathrm{t}$ ) and 2020 (7,629 t) excluded experimental ( E -licence) and out-of-zone catches (O-licence).

## Length analyses

## Length Based Indicators

Yearly 'traffic light' length indicators for females and males are summarized in Tables IV and $V$, respectively. Indicator $L_{c} / L 50$, for conservation of immature fish, were positive (green) in years 2003 and 2013 for females; the rest of the years conservation of immature females was negative (red) or was of concern (yellow). Conservation of immature males was negative or was of concern most years, except for 2013 (green). Indicator $L_{25 \%} / \mathrm{L} 50$, also for conservation of immature fish, showed positive outputs for females at the start of the time series and sporadic positive conservation from 2011 to 2016; most years were of concern. In contrast, conservation of immature males was negative or was of concern most years, including 2019 and 2020. Indicator $\mathrm{L}_{\text {max5 } 5 /} / \mathrm{L}_{\infty}$, for conservation of large individuals, was
negative or of concern throughout the time series for females; this indicator was mostly of concern for males with the exception of 2010 (positive) and 2020 (failed). Indicator $\mathrm{P}_{\text {mega }}$, for the presence of mega-spawners, was all negative for females and mostly negative for males since 2002. Indicator Lmeanc/Lopt, for optimal yield, was negative or of concern for females most years. However, $L_{\text {meanc }} / L_{\text {opt }}$ for males fluctuated between concerning and positive outputs through the time series, and was negative in 2017 and 2018. Indicator $L_{\text {meanc }} / L_{F=M}$, for maximum sustainable yield, was mostly negative for females, with the exception of positive outputs between 2006 and 2015. In contrast, it fluctuated between concerning and positive outputs for males most years, with negative outputs from 2017 to 2020.

Table IV. Female hoki indicators by year, with 'traffic light' scoring. Lc) Length at half the modal catch length; L50) Length at $50 \%$ maturity; $\mathrm{L}_{25 \%}$ ) Length at cumulative $25^{\text {th }}$ percentile of catch; $\mathrm{L}_{\text {max } 5 \%}$ ) Mean length of the largest $5 \%$ of individuals in the catch; $\mathrm{L}_{\infty}$ ) Asymptotic average maximum body size; $\mathrm{P}_{\text {mega }}$ ) Proportion of 'Mega-spawners' in the catch; $\mathrm{L}_{\text {meanc }}$ ) Mean length of individuals larger than LC; Lopt) Optimum length; $L_{F}=M$ ) Length-based proxy for MSY.

|  | Conservation |  |  |  |  | Optimal yield |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ref. | $\mathrm{L}_{\mathrm{c}} / \mathrm{L} 50$ | $\mathrm{~L}_{25 \%} / \mathrm{L} 50$ | $\mathrm{~L}_{\text {max5\% }} / \mathrm{L}_{\infty}$ | $\mathrm{P}_{\text {mega }}$ | $\mathrm{L}_{\text {meanc }} / L_{\text {opt }}$ | $\mathrm{L}_{\text {meanc }} / \mathrm{L}_{\mathrm{F}=\mathrm{M}}$ |
|  | $>1$ | $>1$ | $>0.8$ | $>0.3$ | $\approx 1$ | $\geq 1$ |
| 2002 | 0.99 | 1.08 | 0.67 | 0.00 | 0.78 | 0.92 |
| 2003 | 1.04 | 1.13 | 0.72 | 0.02 | 0.83 | 0.95 |
| 2004 | 0.86 | 0.91 | 0.72 | 0.02 | 0.75 | 0.94 |
| 2005 | 1.00 | 1.04 | 0.76 | 0.04 | 0.86 | 0.96 |
| 2006 | 0.82 | 1.00 | 0.76 | 0.03 | 0.85 | 1.05 |
| 2007 | 0.95 | 1.00 | 0.77 | 0.05 | 0.87 | 0.98 |
| 2008 | 0.91 | 1.00 | 0.77 | 0.04 | 0.85 | 0.98 |
| 2009 | 0.86 | 0.95 | 0.81 | 0.07 | 0.86 | 1.02 |
| 2010 | 0.78 | 1.00 | 0.81 | 0.09 | 0.87 | 1.09 |
| 2011 | 0.92 | 1.05 | 0.81 | 0.11 | 0.90 | 1.04 |
| 2012 | 0.97 | 1.02 | 0.79 | 0.05 | 0.84 | 0.96 |
| 2013 | 1.07 | 1.12 | 0.80 | 0.08 | 0.91 | 0.99 |
| 2014 | 0.89 | 0.99 | 0.69 | 0.01 | 0.74 | 0.91 |
| 2015 | 0.86 | 1.00 | 0.75 | 0.04 | 0.78 | 1.00 |
| 2016 | 0.97 | 1.06 | 0.69 | 0.00 | 0.74 | 0.91 |
| 2017 | 0.69 | 0.74 | 0.64 | 0.00 | 0.54 | 0.80 |
| 2018 | 0.85 | 0.90 | 0.63 | 0.00 | 0.61 | 0.85 |
| 2019 | 1.02 | 1.02 | 0.61 | 0.00 | 0.61 | 0.79 |
| 2020 | 1.04 | 1.04 | 0.60 | 0.00 | 0.63 | 0.81 |

Table V. Male hoki indicators by year, with 'traffic light' scoring. Lc) Length at half the modal catch length; L50) Length at $50 \%$ maturity; $L_{25 \%}$ ) Length at cumulative $25^{\text {th }}$ percentile of catch; $L_{\text {max55\% }}$ ) Mean length of the largest $5 \%$ of individuals in the catch; $L_{\infty}$ ) Asymptotic average maximum body size; $\mathrm{P}_{\text {mega }}$ ) Proportion of 'Mega-spawners' in the catch; Lmeanc) Mean length of individuals larger than LC; Lopt) Optimum length; $L_{F}=M$ ) Length-based proxy for MSY

| Ref. | Conservation |  |  |  | Optimal yield | MSY |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \mathrm{L}_{\mathrm{c}} / \mathrm{L} 50 \\ >1 \end{gathered}$ | $\begin{gathered} \mathrm{L}_{25 \%} / \mathrm{L} 50 \\ >1 \end{gathered}$ | $\begin{gathered} \mathrm{L}_{\max 5 \%} / \mathrm{L}_{\infty} \\ >0.8 \end{gathered}$ | $\begin{aligned} & P_{\text {mega }} \\ & >0.3 \end{aligned}$ | $\begin{gathered} L_{\text {meanc }} / L_{\text {opt }} \\ \quad \approx 1 \end{gathered}$ | $\begin{gathered} L_{\text {meanc }} / L_{F=M} \\ \geq 1 \end{gathered}$ |
| 2002 | 0.92 | 0.96 | 0.76 | 0.03 | 0.91 | 0.97 |
| 2003 | 0.97 | 1.01 | 0.82 | 0.12 | 0.98 | 1.00 |
| 2004 | 0.81 | 0.85 | 0.81 | 0.08 | 0.86 | 0.97 |
| 2005 | 0.94 | 0.98 | 0.82 | 0.15 | 0.97 | 0.99 |
| 2006 | 0.73 | 0.82 | 0.82 | 0.12 | 0.91 | 1.07 |
| 2007 | 0.91 | 0.91 | 0.83 | 0.13 | 0.97 | 1.00 |
| 2008 | 0.87 | 0.92 | 0.80 | 0.08 | 0.93 | 0.98 |
| 2009 | 0.84 | 0.88 | 0.85 | 0.11 | 0.93 | 1.02 |
| 2010 | 0.75 | 0.93 | 0.88 | 0.19 | 0.95 | 1.11 |
| 2011 | 0.85 | 0.94 | 0.86 | 0.22 | 0.95 | 1.06 |
| 2012 | 0.91 | 1.00 | 0.82 | 0.13 | 0.91 | 0.98 |
| 2013 | 1.05 | 1.05 | 0.84 | 0.16 | 0.99 | 0.99 |
| 2014 | 0.88 | 0.93 | 0.74 | 0.02 | 0.83 | 0.94 |
| 2015 | 0.89 | 0.94 | 0.81 | 0.10 | 0.87 | 1.00 |
| 2016 | 0.95 | 1.00 | 0.74 | 0.03 | 0.82 | 0.93 |
| 2017 | 0.67 | 0.72 | 0.66 | 0.00 | 0.59 | 0.82 |
| 2018 | 0.83 | 0.88 | 0.68 | 0.01 | 0.67 | 0.86 |
| 2019 | 0.99 | 0.99 | 0.66 | 0.01 | 0.70 | 0.82 |
| 2020 | 1.01 | 1.01 | 0.62 | 0.00 | 0.72 | 0.84 |

## Length-age relationship

The length-age relationship of females and males pooled for the entire time series ( n $=5,129$ ) gave the following values: $\mathrm{L}_{\infty}=50.80 \mathrm{~cm}, \mathrm{k}=0.1132$, and $\mathrm{t}_{0}=-2.2064$ years. Length and age of females ( $n=3,013$ ) ranged from 12 cm to 47 cm , and from 1 year to 16 years, respectively. The length-age relationship of females gave the following values: $L_{\infty}=53.48 \mathrm{~cm}$, $k=0.1069$, and $\mathrm{t}_{0}=-2.2254$ years. Length and age of males $(\mathrm{n}=2,116)$ ranged from 12 cm to 43 cm and from 1 year to 15 years, respectively. The length-age relationship of males gave the following values: $L_{\infty}=43.10 \mathrm{~cm}, \mathrm{k}=0.1462$, and $\mathrm{t}_{0}=-1.9815$ years (Appendix VII). Yearly von Bertalanffy parameters are summarized in Appendix VIII. Asymptotic lengths ( $\mathrm{L}_{\infty}$ ) of females fluctuated through the time series, with the LOESS smooth between 40 and 70 cm . For males, asymptotic lengths increased significantly from 2010 to 2018 (Fig. 6).


Fig. 6. Asymptotic lengths $\left(L_{\infty}\right)$ calculated according to the von Bertalanffy growth function for female and male hoki, 1988 to 2018. Dark blue lines and light blue areas are the LOESS smooths $\pm 95 \%$ confidence intervals. Yearly data correspond to $L_{\infty}$ in Appendix VIII.

## Length at 50\% maturity (L50)

Lengths at 50\% maturity of females remained relatively stable from 2002 to 2010, and saw a slight decline from 2011 to 2020. Lengths at $50 \%$ maturity of males had a declining trend since the year 2002 but this decrease was steeper since 2011 (Fig. 7).


Fig. 7. Lengths at $50 \%$ maturity (L50) of female and male hoki, 2002 to 2020. Dark blue lines and light blue areas are the LOESS smooths $\pm 95 \%$ confidence intervals. Yearly data correspond to the L50 intercepts in Appendix IX and Appendix X.

## Length frequencies

Female hoki were in the range of sizes from 10 cm to 48 cm pre-anal length, and male individuals were in the range of sizes from 11 cm to 46 cm pre-anal length. Overlap in sizes did not allow certain identification of the total number of cohorts present per year for both, females and males. Individuals $>25 \mathrm{~cm}$ pre-anal length and $\geq 4$ years old were dominant in 2002-2003, and in 2006-2007. Individuals < 25 cm pre-anal length and < 4 years old were
more common in 2004-2005, and from 2007. A peak of nearly 15 cm pre-anal length and 1year old individuals was observed in 2017. The presence of new cohorts was detected in 2004, 2006, 2010, 2013, and 2017, suggesting recruitment to the fishery of 1- and 2-years old individuals, every 2 to 4 years (Fig. 8; Appendix XI).


Fig. 8. Length frequency distribution of female and male hoki in Falkland Islands waters. The progression of sizes of the main cohorts through time are indicated by the dotted lines.

## Conclusions

The indicator of optimal yield was of concern (yellow) or negative (red) mainly for females through the time series. The MSY indicator was negative since 2016 for females and since 2017 for males; in addition, conservation of immature individuals and large individuals was of concern or negative, and conservation of mega-spawners was negative through the time series. These findings are consistent with declines in length at $50 \%$ maturity over the past few years, in particular for males. Length frequency analysis per range of depth, considering the spatial and temporal variability of the presence of hoki may provide greater resolution of length trends. Length Based Indicators suggest that hoki productivity is currently poor; conservation measures should be implemented considering that CPUE has increased substantially in the most recent years. Based on the ICES category 5 assessment framework, a Total Allowable Catch of $6,478 \mathrm{t}$ is recommended for hoki in the year 2022, which represents a decrease of $15 \%$ from the total commercial catch in 2020 (7,629 t).

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## Appendix

Appendix I. Identifying finfish-licenced observer samples.
The FIFD observer database identifies samples by vessel, date, activity (fishing gear type), and observer station, but does not directly link to the licence that the vessel was operating under. If required, the licence must be cross-referenced from the catch report. In most cases, a catch report is recorded the same day by the same vessel, and the corresponding licence can be applied to the samples directly. However, in some cases a catch report is not recorded the same day and instead the nearest catch report by the same vessel either up to 3 days later or 1 day earlier is applied (which still does not result in all samples getting matched). The rationale being that a vessel will file its catch report when it has finished processing the trawl, which may be several days if it is a big haul or the factory is backed up; alternatively, the observer might only sample a trawl the day after it was hauled.

Among positive licence matches, finfish trawl samples are those with activity codes B (bottom trawl), P (pelagic trawl) or S (semi-pelagic trawl), and licence codes $\mathrm{A} / \mathrm{Y}$ (unrestricted finfish), G (Illex + restricted finfish), W/Z (restricted finfish), and S (surimi). Licence code E (experimental) may be any gear or catch target, and can therefore only be matched as finfish by checking against a survey report for that date range or, more expediently, evaluating the species composition that was caught. For this assessment, the criteria were used that a trawl E licence target was designated IIlex if IIlex comprised $>50 \%$ of the catch within 1 day earlier and three days later, skate if skate comprised $>50 \%$ of the catch within 1 day earlier and three days later, and calamari if calamari comprised $>25 \%$ of the catch within 1 day earlier and three days later; otherwise finfish. The lower threshold for calamari reflected the outcome that calamari catch is often scarce in early days of pre-season surveys (e.g., Winter et al. 2019). As criteria of $>50 \%$ Illex / skate vs. $>25 \%$ calamari are non-exclusive, the additional rule was set that a catch composition was designated to that target which exceeded its threshold by the highest proportion. Finfish-designated E licence samples were then added to the commercial licence finfish samples.

Appendix II. Annual commercial catches (t) of hoki reported in Falkland Islands (excluding Elicence; http://www.fig.gov.fk/fisheries/publications/fishery-statistics; Falkland Islands Government 2021),

Argentina (https://www.agroindustria.gob.ar/sitio/areas/pesca_maritima/desembarques/; Sánchez et al. 2012; Navarro et al. 2014, 2019) and Chile (http://www.sernapesca.cl/informes/estadisticas; SERNAPESCA 1990, 2000, 2011, 2021).

| Year | Falkland Islands (t) | Argentina (t) | Chile (t) |
| :---: | :---: | :---: | :---: |
| 1987 | 19,307 | 782 | 131,834 |
| 1988 | 12,209 | 6,952 | 211,624 |
| 1989 | 13,313 | 3,085 | 227,393 |
| 1990 | 7,031 | 4,353 | 128,002 |
| 1991 | 4,499 | 5,583 | 164,697 |
| 1992 | 14,195 | 9,534 | 214,324 |
| 1993 | 8,516 | 29,174 | 82,580 |
| 1994 | 10,055 | 17,472 | 81,310 |
| 1995 | 15,606 | 25,228 | 206,734 |
| 1996 | 13,849 | 46,241 | 375,446 |
| 1997 | 13,020 | 41,787 | 71,479 |
| 1998 | 22,334 | 96,218 | 354,184 |
| 1999 | 18,692 | 118,356 | 309,904 |
| 2000 | 19,846 | 123,926 | 91,333 |
| 2001 | 19,471 | 112,539 | 162,082 |
| 2002 | 26,975 | 98,865 | 133,418 |
| 2003 | 23,764 | 97,797 | 85,896 |
| 2004 | 25,898 | 116,965 | 71,177 |
| 2005 | 16,646 | 115,340 | 79,755 |
| 2006 | 19,425 | 124,638 | 73,421 |
| 2007 | 16,524 | 98,808 | 63,697 |
| 2008 | 15,765 | 110,269 | 73,567 |
| 2009 | 23,219 | 110,717 | 78,440 |
| 2010 | 19,074 | 82,855 | 74,330 |
| 2011 | 22,906 | 70,903 | 70,137 |
| 2012 | 15,815 | 59,595 | 62,175 |
| 2013 | 16,716 | 55,966 | 47,602 |
| 2014 | 7,336 | 58,396 | 39,345 |
| 2015 | 6,782 | 50,469 | 37,475 |
| 2016 | 11,509 | 34,946 | 28,108 |
| 2017 | 3,974 | 21,930 | 20,850 |
| 2018 | 4,408 | 37,598 | 17,055 |
| 2019 | 7,398 | 36,038 | 13,006 |
| 2020 | 7,629 | 31,239 | 12,792 |

Appendix III. Monthly CPUE of hoki in Falkland Islands waters from 1990 to 2020, estimated from G- and W-licensed vessels.


Appendix IV. Monthly CPUE of hoki in Falkland Islands waters during 2020, estimated from G- and W-licensed vessels.


## Appendix IV. continued...



Appendix V. Densities of hoki modelled by inverse distance weighting throughout the Falkland Islands fishing zone, in February 2010-2021.


Appendix V. continued...


Appendix VI. Densities of hoki modelled by inverse distance weighting throughout the Falkland Islands fishing zone, in July 2017 and July 2020.


Appendix VII. von Bertalanffy age-length relationship of female and male hoki from the Falkland Islands.


Appendix VIII. Hoki von Bertalanffy length-at-age parameters for curvature (k), age of fish at length zero ( $t_{0}$ ), and asymptotic length $\left(L_{\infty}\right)$, by year and sex, with $95 \%$ confidence intervals.

| Sex | Year | N |  | k | $\mathrm{t}_{0}$ (years) |  | $\mathrm{L}_{\infty}(\mathrm{cm})$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1988 | 1398 | 0.12 | (0.11-0.13) | -0.55 | (-0.72--0.39) | 56.24 | (54.6-58.2) |
|  | 1989 | 299 | 0.09 | (0.07-0.1) | -2.16 | (-2.76--1.65) | 62.20 | (57.8-68.7) |
|  | 1990 | 241 | 0.08 | (0.04-0.12) | -2.28 | (-3.96--1.15) | 66.80 | (55.1-103.1) |
|  | 1991 | 85 | 1.14 | (0.11-3.6) | 2.39 | (-11.88-4.16) | 32.11 | (31.6-36.4) |
|  | 1992 | 181 | 0.13 | (0.08-0.18) | -2.39 | (-3.53--1.59) | 49.91 | (44.8-60.3) |
|  | 1993 | 285 | 0.10 | (0.07-0.13) | -2.40 | (-3.5--1.59) | 54.57 | (49.2-64.5) |
|  | 1994 | 437 | 0.10 | (0.08-0.12) | -1.27 | (-1.87--0.77) | 59.10 | (55.2-64.6) |
|  | 1995 | 366 | 0.04 | (0.01-0.09) | -4.05 | (-7.18--1.81) | 85.89 | (59.7-320.5) |
|  | 1996 | 447 | 0.08 | (0.05-0.1) | -2.41 | (-3.3--1.72) | 62.49 | (55.8-74.4) |
|  | 1997 | 391 | 0.11 | (0.09-0.12) | -0.78 | $(-1.12-0.48)$ | 57.16 | (54-61.4) |
|  | 1998 | 263 | 0.13 | (0.11-0.16) | -0.37 | (-0.8--0.03) | 54.08 | (50.7-58.9) |
|  | 1999 | 384 | 0.08 | (0.06-0.11) | -1.97 | (-2.79--1.31) | 64.81 | (58-76.9) |
|  | 2000 | 151 | 0.11 | (0.08-0.14) | -1.60 | $(-2.36-1)$ | 55.22 | (50.8-62.2) |
|  | 2001 | 401 | 0.04 | (0.02-0.06) | -4.78 | (-6.08--3.81) | 91.57 | (70.3-164.7) |
|  | 2002 | 407 | 0.01 | (0-0.05) | -5.86 | (-6.74--4.13) | 213.07 | (79.8-1021.3) |
| F | 2003 | 298 | 0.13 | (0.1-0.15) | -2.19 | (-2.68--1.79) | 49.47 | (45.8-54.8) |
|  | 2004 | 354 | 0.09 | (0.05-0.13) | -3.05 | (-4.1--2.27) | 54.35 | (46.1-73.6) |
|  | 2005 | 292 | 0.14 | (0.09-0.2) | -1.35 | (-2.45--0.58) | 49.73 | (45-59.4) |
|  | 2006 | 258 | 0.12 | (0.06-0.17) | -1.66 | (-2.56--1.03) | 53.74 | (45.4-75.8) |
|  | 2007 | 511 | 0.11 | (0.08-0.14) | -2.56 | (-3.25--1.99) | 50.79 | (45.8-58.7) |
|  | 2008 | 454 | 0.17 | (0.13-0.22) | -1.07 | (-1.7--0.58) | 45.12 | (41.6-50.8) |
|  | 2009 | 499 | 0.17 | (0.14-0.19) | -0.88 | (-1.2--0.6) | 44.58 | (42.4-47.3) |
|  | 2010 | 392 | 0.17 | (0.13-0.21) | -0.89 | $(-1.3--0.56)$ | 43.57 | (40.5-47.9) |
|  | 2011 | 243 | 0.19 | (0.16-0.23) | -1.08 | (-1.55--0.7) | 45.53 | (43.3-48.6) |
|  | 2012 | 0 | NA | (NA - NA) | NA | (NA - NA) | NA | (NA - NA) |
|  | 2013 | 343 | 0.12 | (0.08-0.15) | -1.99 | (-2.69--1.45) | 53.46 | (48.3-62.5) |
|  | 2014 | 237 | 0.14 | (0.08-0.19) | -1.53 | (-2.31--0.99) | 49.50 | (42.6-65.5) |
|  | 2015 | 383 | 0.13 | (0.09-0.16) | -1.19 | (-1.7--0.78) | 53.20 | (48-61.9) |
|  | 2016 | 371 | 0.14 | (0.1-0.18) | -1.10 | (-1.65--0.66) | 51.68 | (46.2-61.2) |
|  | 2017 | 278 | 0.09 | (0.06-0.12) | -1.75 | (-2.28--1.33) | 59.16 | (50.9-75.7) |
|  | 2018 | 262 | 0.12 | (0.08-0.16) | -1.70 | (-2.46--1.14) | 53.05 | (46.8-65.6) |


| Sex | Year | N | k | $\mathrm{t}_{0}$ (years) |  | $\mathrm{L}_{\infty}(\mathrm{cm})$ |  |
| :--- | ---: | ---: | ---: | :--- | ---: | :--- | :--- | :--- |
| 1988 | 1031 | 0.20 | $(0.18-0.22)$ | 0.15 | $(-0.02-0.33)$ | 44.82 | $(43.4-46.4)$ |
| 1989 | 199 | 0.06 | $(0.04-0.09)$ | -3.16 | $(-4.29--2.32)$ | 69.74 | $(58.5-97.1)$ |
| 1990 | 141 | 0.08 | $(0.03-0.14)$ | -2.51 | $(-4.69--1.23)$ | 60.55 | $(49.5-108.2)$ |
| 1991 | 59 | 0.57 | $(0.09-1.61)$ | 0.90 | $(-11.89-3.41)$ | 32.97 | $(31.9-39)$ |
| 1992 | 166 | 0.20 | $(0.15-0.26)$ | -1.63 | $(-2.46--1.01)$ | 41.62 | $(39.2-45.4)$ |
| 1993 | 292 | 0.15 | $(0.1-0.2)$ | -2.06 | $(-3.32--1.2)$ | 43.81 | $(40.4-50.5)$ |
| 1994 | 133 | 0.15 | $(0.1-0.2)$ | -0.88 | $(-2.07--0.09)$ | 45.91 | $(41.9-54)$ |
| 1995 | 201 | 0.06 | $(0.01-0.13)$ | -4.20 | $(-8.48--1.44)$ | 65.19 | $(47-268.9)$ |
| 1996 | 390 | 0.12 | $(0.08-0.15)$ | -1.76 | $(-2.76--1.01)$ | 49.40 | $(45.2-56.9)$ |
| 1997 | 232 | 0.13 | $(0.11-0.16)$ | -0.49 | $(-0.83--0.19)$ | 51.57 | $(47.5-57.3)$ |
| 1998 | 135 | 0.20 | $(0.14-0.28)$ | -0.03 | $(-0.63-0.46)$ | 42.89 | $(38.6-49.2)$ |
| 1999 | 194 | 0.09 | $(0.04-0.14)$ | -2.13 | $(-4.13--1.03)$ | 58.06 | $(48.8-95.9)$ |
| 2000 | 65 | 0.11 | $(0.04-0.18)$ | -1.49 | $(-3.64--0.27)$ | 54.54 | $(45.5-91.6)$ |
| 2001 | 258 | 0.04 | $(0.02-0.07)$ | -6.14 | $(-8.16--4.69)$ | 72.85 | $(55.7-139.9)$ |

3 | 2002 | 279 | 0.10 | $(0.06-0.14)$ | -3.11 | $(-4.38--2.21)$ | 49.63 | $(42.9-65.1)$ |
| ---: | ---: | ---: | ---: | :--- | ---: | :--- | ---: | :--- |
| 2003 | 198 | 0.18 | $(0.14-0.22)$ | -1.64 | $(-2.16--1.22)$ | 40.56 | $(37.6-45)$ |
| 2004 | 285 | 0.16 | $(0.09-0.24)$ | -2.76 | $(-4.24--1.8)$ | 38.97 | $(34.8-49.3)$ |
| 2005 | 67 | 0.16 | $(0.1-0.23)$ | -1.13 | $(-2.24--0.38)$ | 46.54 | $(41-58.1)$ |
| 2006 | 144 | 0.22 | $(0.1-0.34)$ | -1.06 | $(-2.38--0.32)$ | 39.32 | $(34.2-55.4)$ |
| 2007 | 414 | 0.14 | $(0.1-0.18)$ | -1.82 | $(-2.46-1.32)$ | 44.50 | $(40.3-51.6)$ |
| 2008 | 286 | 0.27 | $(0.18-0.37)$ | -0.71 | $(-1.52--0.14)$ | 35.49 | $(32.8-40.2)$ |
| 2009 | 360 | 0.19 | $(0.15-0.23)$ | -0.89 | $(-1.34--0.52)$ | 40.34 | $(37.7-44)$ |
| 2010 | 260 | 0.20 | $(0.15-0.25)$ | -0.98 | $(-1.5--0.55)$ | 38.73 | $(35.9-43.1)$ |
| 2011 | 130 | 0.17 | $(0.11-0.23)$ | -1.79 | $(-2.69-1.15)$ | 42.92 | $(38.6-50.7)$ |
| 2012 | 0 | NA | $(N A-N A)$ | NA | $(N A-N A)$ | NA | $(N A-N A)$ |
| 2013 | 177 | 0.12 | $(0.07-0.17)$ | -1.84 | $(-2.67--1.24)$ | 51.18 | $(44-66.8)$ |
| 2014 | 167 | 0.08 | $(0.01-0.16)$ | -2.84 | $(-4.39--1.69)$ | 62.18 | $(42-260.6)$ |
| 2015 | 294 | 0.08 | $(0.02-0.13)$ | -1.98 | $(-3.12--1.21)$ | 68.92 | $(50.3-178)$ |
| 2016 | 252 | 0.17 | $(0.11-0.23)$ | -1.09 | $(-1.76--0.57)$ | 43.30 | $(38.5-52)$ |
| 2017 | 226 | 0.04 | $(0.01-0.09)$ | -1.88 | $(-2.39--1.31)$ | 119.04 | $(62.5-691.4)$ |
| 2018 | 177 | 0.05 | $(0.01-0.11)$ | -3.00 | $(-4.4--1.86)$ | 81.93 | $(52.6-371.8)$ |

Appendix IX. Binomial logistic regressions of juvenile (0) or adult (1) maturity vs. length for female hoki. Red lines: Length intercept of $50 \%$ adulthood, corresponding to Fig. 7.


Appendix X. Binomial logistic regressions of juvenile (0) or adult (1) maturity vs. length for male hoki. Red lines: Length intercept of $50 \%$ adulthood, corresponding to Fig. 7.


Appendix XI. Number of hoki individuals sampled for length frequency distributions.

| Year | Females (n) | Males (n) |
| :--- | :---: | :---: |
| 2002 | 11,626 | 8,650 |
| 2003 | 3,713 | 2,608 |
| 2004 | 5,408 | 4,181 |
| 2005 | 6,000 | 4,378 |
| 2006 | 3,383 | 2,243 |
| 2007 | 4,563 | 3,158 |
| 2008 | 4,449 | 3,090 |
| 2009 | 9,677 | 7,476 |
| 2010 | 2,875 | 2,058 |
| 2011 | 1,503 | 1,135 |
| 2012 | 1,957 | 1,289 |
| 2013 | 2,749 | 1,737 |
| 2014 | 1,460 | 905 |
| 2015 | 746 | 588 |
| 2016 | 3,213 | 1,713 |
| 2017 | 1,273 | 1,253 |
| 2018 | 1,814 | 1,475 |
| 2019 | 1,372 | 996 |
| 2020 | 1,514 | 1,076 |


[^0]:    ${ }^{\text {a }}$ An additional one-day transect of four trawls was taken in shallow inshore waters to sample for juvenile toothfish. These four trawls were not included in analyses as their locations were not relevant to the distribution of hoki.

